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Isotopic compositions of ^{236}U and Pu isotopes in “Black Substances” collected from roadsides in Fukushima Prefecture: fallout from the Fukushima Dai-ichi Nuclear Power Plant accident

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1 **Isotopic compositions of ^{236}U and Pu isotopes in “Black Substances” collected from**
2 **roadsides in Fukushima Prefecture: fallout from the Fukushima Dai-ichi Nuclear Power**
3 **Plant accident**

4
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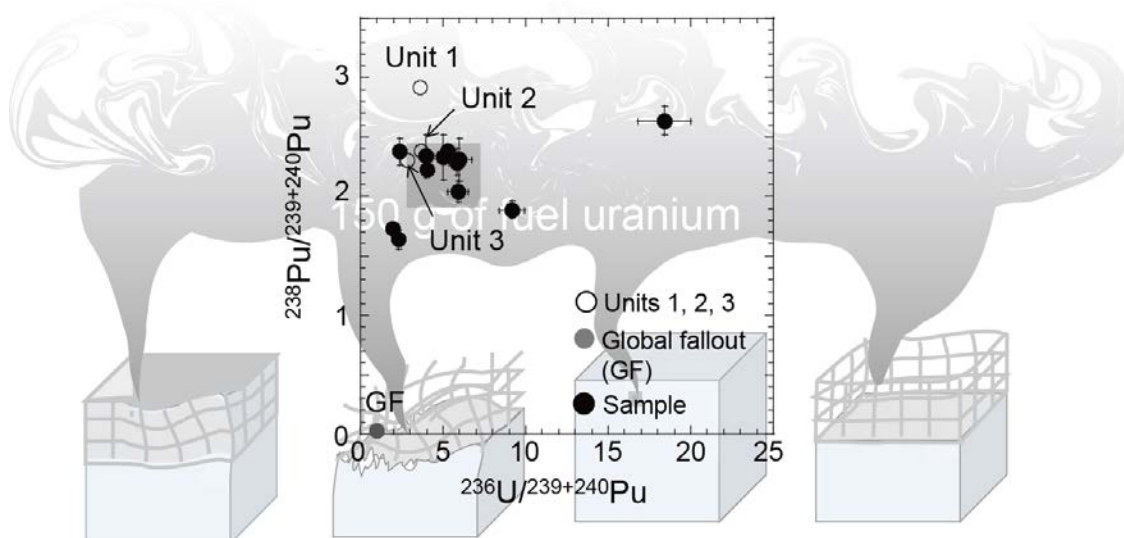
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15 **ABSTRACT**

16 Black colored road dusts were collected in high radiation areas in Fukushima Prefecture.
17 Measurement of ^{236}U and Pu isotopes and $^{134,137}\text{Cs}$ in samples was performed in order to confirm
18 whether refractory elements such as U and Pu from the fuel core were discharged, and to
19 ascertain the extent of fractionation between volatile and refractory elements. The
20 concentrations of $^{134,137}\text{Cs}$ in all samples were exceptionally high, ranging from 0.43 to 17.7
21 MBq/kg, respectively. $^{239+240}\text{Pu}$ was detected at low levels, ranging from 0.18 to 1.14 Bq/kg and
22 with high $^{238}\text{Pu}/^{239+240}\text{Pu}$ activity ratios of 1.64 to 2.64. ^{236}U was successfully determined in the
23 range 0.28 to 6.74×10^{-4} Bq/kg. The observed activity ratios for $^{236}\text{U}/^{239+240}\text{Pu}$ were in reasonable
24 agreement with those calculated for the fuel core inventories, indicating that trace amounts of U
25 from the fuel cores were released together with Pu isotopes, but without large fractionation. The
26 quantities of U and $^{239+240}\text{Pu}$, emitted to the atmosphere were estimated as 2.3×10^9 Bq (150 g)
27 and 3.9×10^6 Bq (580 mg), respectively. Regarding U, this is the first report to give a quantitative
28 estimation of the amount discharged. Appreciable fractionation between volatile and refractory
29 radionuclides associated with the dispersal/deposition processes with distance from the
30 Fukushima Dai-ichi Nuclear Power Plant was found.



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INTRODUCTION

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In the Fukushima Dai-ichi Nuclear Power Plant (FDNPP) accident, which occurred following the Great East Japan Earthquake of March 11, 2011, large amounts of radionuclides, especially volatile species such as ^{131}I ($T_{1/2}=8.02$ d), ^{134}Cs ($T_{1/2}=2.06$ y) and ^{137}Cs ($T_{1/2}=30.2$ y), were released into the environment mainly from Units 1-3¹ of the plant. A possibility of radionuclide emission from the spent fuel pool of Unit 4, which had been shut down since November 2010, has been discounted due to removal of fuel lods which commenced in November 2013. That is, damage to the spent fuel rods has not been serious and there has been minimal contribution of nuclide emissions from this source. Given this situation, there has been some public concern as to whether U and Pu isotopes in fuel were discharged to the environment, because some mixed U and Pu oxide fuel (MOX fuel) was used in Unit 3, and partial melt-down of the nuclear fuel cores occurred in Units 1-3. Due to the difficulty of measuring U and Pu radionuclides, data on the concentrations and the isotopic composition of U and Pu in the contaminated areas of Fukushima have been limited²⁻⁷. Zheng et al. (2012, 2013)^{5,6} reported isotopic evidence for the release of Pu based on $^{241}\text{Pu}/^{239+240}\text{Pu}$ activity ratios in surface soil samples collected from the 30 km zone of FDNPP, where values were more than 100 compared with the global fallout value of about 1.2 in March 2011. Yamamoto et al. (2012, 2014)^{4,7} also reported $^{238}\text{Pu}/^{239+240}\text{Pu}$ activity ratios one to two orders of magnitude higher than the global fallout value of ca. 0.03, found in surface soils from surrounding villages and towns. In fact, about a few % to 50% of $^{239+240}\text{Pu}$ found in these samples was estimated to have originated from the FDNPP accident, based on the activities and/or atom ratios of Pu isotopes. As for U released from the fuel cores, ^{235}U and ^{238}U

55 have been measured on soil samples collected around the contaminated areas, but no anomaly
56 for ^{235}U associated with FDNPP fallout was observed because of the presence of a much larger
57 quantity of natural U in soil². Uranium-236 ($T_{1/2}=2.342\times 10^7$ y) is mainly produced via the
58 nuclear reaction $^{235}\text{U}(n,\gamma)$ by thermal neutrons in reactors which use ^{235}U fuel. Unambiguous
59 detection of ^{236}U from the FDNPP accident, however, has not yet been achieved. Although
60 Sakaguchi et al. (2012)³ measured this nuclide in seawater samples, an assessment of the source
61 was not possible. Thus, the characteristics of U and transuranic elements as refractory elements
62 and emitted from the FDNPP have not been fully characterised in the environment, even two
63 years after the accident.

64 Following the accident, road dust, blackish in color, has attracted the attention of citizens in
65 the Fukushima and Tokyo areas, the dust being given the name "Black Substances". The
66 material is composed of aerosol particles, asphalt and minute tire particles originating from
67 traffic, residue of dried lichens, soil and other fine-grained environmental debris. The dust,
68 which is blown by wind and rain into street corners and dips in the road, contains extremely
69 high levels of radionuclides. Since there is less of a contribution of natural U and global fallout
70 Pu in the matrix of these samples because of a reduced contribution of soil-derived materials,
71 analysis of the road dust would seem appropriate for gaining new information on the isotopic
72 compositions of U and transuranic elements released from the FDNPP.

73 In this paper, we report the concentration and isotopic compositions of ^{236}U and Pu isotopes
74 (half-lives: ^{238}Pu 87.74 y; ^{239}Pu 2.411×10^4 y; ^{240}Pu 6.563×10^3 y), and $^{134,137}\text{Cs}$, in Black
75 Substances from roadsides which are heavily contaminated by fallout from the FDNPP accident.
76 The aims of research are (1) to gain new knowledge on the levels and isotopic compositions of
77 the radionuclides released, (2) to compare values with those of nuclear fuel cores and further (3)
78 to evaluate the fractionation effects between the volatile and refractory radionuclides which
79 were released into the environment. Some focus is given to the measurement of ^{236}U to clarify
80 whether U was released from the fuel cores. The present study serves as an aid not only to
81 predict the fate of U and Pu in the environment but also to estimate, accurately and precisely,
82 their inventories in the cores of the reactors.

83

84 MATERIAL AND METHODS

85 Sampling of road dust: "Black Substances"

86 The road dust samples were mainly collected from Minami-soma City and Namie Town from
87 May to September, 2012 (Table 1 and Fig. 1). These regions were seriously contaminated by
88 discharge of the radioactive plume due to the venting operations, hydrogen gas detonations and

89 breaches of the containment vessel on March 12-15, 2011, followed by scavenging of the
90 nuclides in the atmosphere by precipitation in the form of rain and snow. The sampling points
91 were located in the evacuation zone and show a relatively high radiation dose ($>19 \mu\text{Sv/h}$) even
92 one year after the accident⁹. The collected substances were air-dried, and sieved through a 2-mm
93 mesh to remove pebbles and plant debris, then pulverized in an agate-mortar for a few hours to
94 obtain homogeneous samples. The homogeneity of sample has been confirmed in previous
95 papers^{7,10}.

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98 **Measurement of radionuclides**

99 **¹³⁷Cs and Pu isotope**

100 Each sample (2-5g) was first packed into a plastic vessel of diameter 4 cm and height of 2 cm.
101 Then, ¹³⁴Cs and ¹³⁷Cs were determined by γ -ray spectrometry using a Ge-detector (Princeton
102 Gamma Tech, relative efficiency 40%) with measurement times of 600-1200 sec for each
103 sample. The spectrometer was calibrated with a standard of the New Brunswick Laboratory
104 (NBL) reference sample No. 42-1 (4.04% uranium), analytical grade KCl and a standard γ -ray
105 reference sample for nuclides including ¹³⁷Cs from the JRIA (Japan Radioisotope Association).
106 In addition, the ¹³⁴Cs concentration was corrected for cascade summing. The detection
107 efficiency for ¹³⁷Cs (662 keV) was 5%, and the detection limit was 4.20 Bq for a 600 s
108 measurement.

109 After γ -ray spectrometric analysis, plutonium analysis was carried out radiochemically^{9,10}. In
110 brief, an aliquot of 30-50 g of the sample was calcinated at about 450°C overnight and Pu was
111 leached twice with 250 ml of concentrated HNO₃ with a small amount of H₂O₂ on a hot plate,
112 with the addition of a known amount of ²⁴²Pu as a yield tracer. Plutonium was then separated
113 from the main matrix and purified carefully by passing through an anion exchange resin column
114 (diameter 1 cm x 10 cm height; Dowex 1-X8, 100-200 mesh). The purified Pu was electroplated
115 onto a polished stainless steel disc, and activities (²³⁸Pu, ²³⁹⁺²⁴⁰Pu and ²⁴²Pu) were measured by
116 α - spectrometry. As the α -particle energies of ²³⁹Pu and ²⁴⁰Pu are nearly the same, these
117 nuclides could not be differentiated. Their activities were, therefore, presented as the sum
118 (²³⁹⁺²⁴⁰Pu) of ²³⁹Pu and ²⁴⁰Pu. The detection limit for Pu was 4.82×10^{-5} Bq for a 1 week
119 measurement.

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121 **U isotope measurements**

122 A portion (3-5 g) of the sample was decomposed with concentrated HF+HNO₃+HClO₄ (10 ml

123 soln. per 1 g sample) for 12 hours on a hot plate (150°C) with continual agitation. The total
124 amount of solution leached was weighed and an aliquot was sub-sampled for determination by
125 ICP-MS, of the total amount of ^{238}U leached. The remaining solution was used to determine
126 the $^{236}\text{U}/^{238}\text{U}$ atom ratio by accelerator mass spectrometry (AMS).

127 Purification of U was conducted as described by Sakaguchi et al. (2009, 2010)^{13,14}. For
128 measurement of $^{236}\text{U}/^{238}\text{U}$ with AMS, the cathode was prepared as U oxides in a Fe_2O_3 matrix.
129 Details of the measurements of ^{236}U with AMS are described in Steier et al. (2010)¹⁵ and
130 Sakaguchi et al. (2010)¹⁴. The detection limit is about 10^4 atoms of ^{236}U . ^{238}U was measured
131 using ICP-MS (Agilent 7700) after appropriate dilution with 2% HNO_3 . In this case, indium was
132 used as an internal standard.

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135 **RESULTS AND DISCUSSION**

136 **Radionuclide composition of Black Substances**

137 The results of the Cs, U and Pu isotope measurements are shown in Tables 2 and 3. The
138 concentrations of ^{134}Cs and ^{137}Cs in all samples were exceptionally high, ranging from 0.43
139 to 11.4 MBq/kg and 0.58 to 17.7 MBq/kg, respectively. These concentrations are two to three
140 orders of magnitude higher than those of surface soil collected from the areas within the 20 km
141 exclusion zone. These extremely high concentrations of radio Cs reflect a relative absence of
142 soil matrix components and the fact that the sampling area where is heavily contaminated with
143 radio Cs even at distance well removed from the FDNPP. Yoshida and Takahashi (2012)¹⁶ have
144 reviewed the reasons why a high radiation area exists far removed from FDNPP, and it is mainly
145 due to the emission of highly radioactive plume combined with wet deposition associated with
146 rain on the 15th of March. The $^{134}\text{Cs}/^{137}\text{Cs}$ activity ratios covered a narrow range of 0.60 to 0.76,
147 depending mainly on the sampling date, and these values approached 1 when they were
148 decay-corrected to the date on 11 March 2011 as reported in METI (2011) and other
149 papers^{1,4,7,10,17,18}.

150 $^{239+240}\text{Pu}$ was detected at low levels in all the samples examined, ranging from 0.18 to 1.14
151 Bq/kg with $^{238}\text{Pu}/^{239+240}\text{Pu}$ activity ratios of 1.64 to 2.64. However, the concentrations of
152 $^{239+240}\text{Pu}$ are higher than those of global fallout observed in Fukushima. Actually, the typical
153 concentration of global fallout $^{239+240}\text{Pu}$ in surface soil of Fukushima^{4,7}, the samples being
154 obtained from undisturbed and non-forested open flat land, is one to two orders of magnitude
155 lower than other areas in Japan (about 1 Bq/kg)^{11,14,19}.

156 The ^{238}Pu activity concentrations in the Black Substances are well correlated with that of

157 $^{239+240}\text{Pu}$ ($R=0.99$), and their correlation slope, that is the ratio, is two orders of magnitude
158 higher than the global fallout value of 0.03. These values agree well with those (2.0 ± 0.5)
159 identified in several surface soil samples taken within the FDNPP site²⁰ and those calculated for
160 the cores of Units 1-3 by the Japan Atomic Energy Agency ($2.30\text{-}2.92$)²¹. Clearly, it is very
161 important to point out that almost all Pu in the black dust samples was originates from the
162 FDNPP accident.

163 ^{236}U was successfully measured in the range of $(0.28\text{ to }6.74)\times 10^{-4}$ Bq/kg. The $^{236}\text{U}/^{238}\text{U}$
164 atom ratios were in the range of $(0.25\text{ to }2.60)\times 10^{-7}$. For spent nuclear fuel, ratios above 10^{-3} are
165 typical. The uranium in Black Substances was obviously diluted with natural ^{238}U in the
166 environment. The atom ratios are comparable with those in soil samples collected from Mogilev,
167 located more than 200 km north of the Chernobyl Nuclear Power Plant²². But the usefulness of
168 this comparison is limited because of the difference in the ^{238}U content of soil and Black
169 Substances. The observed ratios are clearly one to two orders of magnitude higher than those of
170 global fallout values found in surface soil in Japan^{13,14}. The activity ratios of $^{236}\text{U}/^{239+240}\text{Pu}$ were
171 in the range of $(1.96\text{ to }18.4)\times 10^{-4}$ with a weighted mean value of 7.87×10^{-4} . This average value
172 is about 7 times higher than that for the global fallout value for Japan^{13,14}.

173 Figure 2 shows a correlation of the concentration of the fuel burning products ^{238}Pu and
174 ^{236}U in Black Substances. The concentrations of ^{238}Pu and ^{236}U are linearly well correlated. This
175 implies that the deposited nuclides were well homogenized in the transportation process, or
176 there is mainly one or two sources. Figure 3 shows the relationship between the $^{238}\text{Pu}/^{239+240}\text{Pu}$
177 and the $^{236}\text{U}/^{239+240}\text{Pu}$ activity ratios. The global fallout value is also plotted on this figure.
178 Sample no. 1, which was collected from close to the FDNPP (3.43 km), is far from the other
179 samples. The grey rectangle in Fig. 3 gives the one sigma (standard deviation) area of
180 $^{238}\text{Pu}/^{239+240}\text{Pu}$ and $^{236}\text{U}/^{239+240}\text{Pu}$ (except for sample no. 1). By taking into consideration fact that
181 the Pu in these samples was predominantly derived from the FDNPP, the ratios seem to
182 represent the ratios for ^{236}U and ^{238}Pu released into the environment as a results of nuclear
183 accident. The JAEA²¹ reported core inventories for each radionuclide calculated mainly by
184 using the ORIGEN2 code. By referring to those data, the $^{236}\text{U}/^{239+240}\text{Pu}$ activity ratios observed
185 in Black Substances show a relatively good correlation with the estimated core ratios: 3.63×10^{-4} ,
186 3.66×10^{-4} and 2.84×10^{-4} (activity ratio) in Units 1, 2 and 3, respectively (Fig. 3). If we pay
187 attention to the details, the isotopic composition of U and Pu can be explained by mixing of core
188 materials and global fallout for nos. 9, 10 and 11. The extra sources which has (1) high
189 $^{238}\text{Pu}/^{239+240}\text{Pu}$ and high $^{236}\text{U}/^{239+240}\text{Pu}$ ratios and/or (2) low $^{238}\text{Pu}/^{239+240}\text{Pu}$ ratio but high
190 $^{236}\text{U}/^{239+240}\text{Pu}$ ratio such as a low reprocessing fuel with low or no burn up, can be predicted

191 from this graph (broken lines 1 and 2). However, confirmation of this hypothesis has not been
192 found in this study. In any case, it is reasonable to conclude that trace amounts of uranium from
193 the fuel core was released into the environment together with Pu isotopes without large
194 fractionation.

195

196 **Calculation of the amount of Pu and U released**

197 Using the averaged $^{137}\text{Cs}/^{239+240}\text{Pu}$ activity ratio observed in this study and the amount of
198 ^{137}Cs released to the atmosphere (1.5×10^{16} Bq)¹, the amount of $^{239+240}\text{Pu}$ released can be
199 estimated as 2.3×10^9 Bq. This value is smaller than that estimated by METI (6.4×10^9 Bq)¹. The
200 amount of $^{239+240}\text{Pu}$ dispersed to the atmosphere is $3.8 \times 10^{-5}\%$ of the total amount in the cores of
201 reactors 1-3²¹. In fact, this equates to about 590 mg of total plutonium (^{238}Pu , ^{239}Pu , ^{240}Pu). Our
202 estimated values for the amount of $^{239+240}\text{Pu}$ released from the core and the ratio of the released
203 $^{239+240}\text{Pu}$ to the total amount in the core are close to the range of calculated values (1.0 to
204 2.4×10^9 Bq and 1.2 to $2.9 \times 10^{-5}\%$) by Zheng et al. (2013)⁶, respectively.

205 In the same way, the amount of ^{236}U released as a result of the accident was estimated
206 using the observed $^{236}\text{U}/^{239+240}\text{Pu}$ and $^{236}\text{U}/^{137}\text{Cs}$ activity ratio. The amount of ^{236}U released to the
207 atmosphere was about 1.2×10^6 Bq. This is about $6.1 \times 10^{-5}\%$ of the ^{236}U inventories in cores 1-3
208 (2.0×10^{12} Bq)²¹. This release ratio possibly applies also to the other uranium isotopes such as
209 ^{234}U , ^{235}U and ^{238}U . From the uranium inventories in the cores 1-3²¹, the total amount of uranium
210 isotopes, ^{234}U , ^{235}U , ^{236}U and ^{238}U , dispersed in the atmosphere has been estimated to be about
211 3.9×10^6 Bq. This equates to about 150 g of total uranium, which is four to five orders of
212 magnitude smaller than that of Chernobyl²³⁻²⁶. In the FDNPP, the enrichment ratio of ^{235}U is
213 about 1.7 to 1.9% for each reactor, and is obviously different from that of natural ^{235}U (0.72%).
214 However, relatively high concentrations of natural uranium in Japanese soil (about 1-3 ppm)
215 prevent identification of the ^{235}U fingerprint of uranium contamination from the FDNPP
216 especially for the granitoid area in Fukushima. In this situation, ^{236}U is very appropriate for
217 demonstrating anthropogenic uranium contamination in the surface environment. Summary for
218 the calculation of the amount of Pu and U released is shown in Table 5.

219

220 **Estimation of fractionation factor between refractory and volatile elements**

221 It is of interest to evaluate the fractionation factor between refractory (R) and volatile (V)
222 elements. The R/V factor is often used to predict the fate of radionuclides in the environment²⁶
223 and is an essential parameter to estimate radionuclide dispersion as a result of the venting
224 operations, hydrogen detonations and breaches in containment. Here, three categories of R/V

225 factors were calculated based on the results for ^{137}Cs , $^{239+240}\text{Pu}$, ^{238}Pu and ^{236}U concentrations.
226 The R/V factor are defined as:

227

228 R_{Pu}/V factor = (Measured ($^{239+240}\text{Pu}/^{137}\text{Cs}$) in the environmental sample)/(inventory ratio of
229 ($^{239+240}\text{Pu}/^{137}\text{Cs}$) in the reactor)

230 R_{Pu8}/V factor = (Measured ($^{238}\text{Pu}/^{137}\text{Cs}$) in the environmental sample)/(inventory ratio of
231 ($^{238}\text{Pu}/^{137}\text{Cs}$) in the reactor)

232 R_{U}/V factor = (Measured ($^{236}\text{U}/^{137}\text{Cs}$) in the environmental sample)/(inventory ratio of
233 ($^{236}\text{U}/^{137}\text{Cs}$) in the reactor)

234

235 $^{239+240}\text{Pu}/^{137}\text{Cs}$ (Bq/Bq), $^{238}\text{Pu}/^{137}\text{Cs}$ (Bq/Bq) and $^{236}\text{U}/^{137}\text{Cs}$ (Bq/Bq) in the core of Units 1, 2
236 and 3 were estimated using the calculation of Nishihara et al. (2012)²¹ (Table 4). From the
237 averages of these calculated values ($^{239+240}\text{Pu}/^{137}\text{Cs}=8.54\times 10^{-3}$, $^{238}\text{Pu}/^{137}\text{Cs}=2.12\times 10^{-2}$ and
238 $^{236}\text{U}/^{137}\text{Cs}=2.78\times 10^{-6}$) and our measured data for each sample, the R/V factors were estimated to
239 be nearly the same and in the range of $(0.33\text{ to }4.49)\times 10^{-5}$ for R_{Pu}/V , $(0.18\text{ to }1.76)\times 10^{-5}$ for
240 R_{Pu8}/V and $(0.38\text{ to }8.21)\times 10^{-5}$ for R_{U}/V . These R/V values are of the same order of magnitude
241 as data based on $^{239+240}\text{Pu}/^{137}\text{Cs}$ found for surface soil/water samples contaminated with Pu
242 derived from the FDNPP^{3,4,7}. These R_{Pu}/V , R_{Pu8}/V and R_{U}/V values are clearly small, confirming
243 that a much larger amount of volatile nuclides such as Cs and I have been dispersed and
244 introduced into the surface environment compared to refractory elements such as U and Pu.

245 Surprisingly, the R/V factors, especially for samples with U and Pu isotopic compositions
246 inside the gray rectangle in Fig. 3, increased with distance from the FDNPP (Fig. 4). This
247 behavior was unexpected, since the refractory elements in Chernobyl were emitted as larger
248 particles, and consequently should have been deposited more quickly than the volatile elements,
249 which were emitted as gases and which accumulated on ambient aerosol particles with a typical
250 size of 0.1 to 1 μm (accumulation or droplet mode²⁷). Here in our results, different fractionation
251 mechanism may operate to induce a preferential deposition of refractory elements at distances
252 further from the FDNPP. However, the gradient of the regression curve for R_{Pu8}/V is different
253 from that of $R_{\text{Pu}}(R_{\text{U}})/V$. For two isotopes of the same element, ^{238}Pu and $^{239+240}\text{Pu}$, fractionation
254 in the deposition process is no valid explanation. In addition as shown in Fig. 2, fractionation
255 between ^{236}U and ^{238}Pu with the distance from the FDNPP is also not evident. Thus, it is
256 expected that the increase of this R/V factor with distance is governed by a different source.

257 However, from these preliminary results it would be premature to draw firm conclusions
258 on the relationships observed in Fig. 4. Inclusion of more sampling sites, a higher

259 time-resolution for the release history, different behaviors of the radionuclides after deposition
260 and precipitation data could easily change the picture. In fact, no clear correlation between R/V
261 factors, or with distance or precipitation has been found for a much larger data set for
262 $^{239+240}\text{Pu}/^{137}\text{Cs}$ in about 150 environmental samples including surface soil and Black
263 Substance^{7,10}.

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270 Notes

271 The authors declare no competing financial interest.

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Table 1 Sampling locations of soil samples (Black Substances) including dust etc. collected mainly from the roadside

Sample No.	Lab ID	Sampling date	Position (GPS data)		distance from FDNPP (km)	Location	Remarks
			N	E			
1	12N-7	22-Sep-12	37°27'12.8"	141° 00'40.6"	3.43	Futaba	Side of the parking area at Futabakousei hospital
2	12N-17	22-Sep-12	37°28'02.4"	140° 55'51.0"	10.2	Namie	Roadsides (Route 253) in front of Matsunaga pottery
3	12N-23	22-Sep-12	37°28'14.6"	140° 55'59.2"	10.19	Namie	Space under the eaves of Yasuragi nursing home
4	12N-30	22-Sep-12	37°30'25.8"	140° 55'32.6"	13.11	Namie	Rooftop of a concrete house in Murohara (Route 114)
5	12R-7	23-Apr-12	37°32'59.6"	140° 56' 04.3"	16.16	Minami-soma	Side of a road near Odaka-chiku (Kanaya)
6	KURO-3	27-Mar-12	37°35'59.5"	140° 56' 21.5"	21.07	Minami-soma	Vicinity of the bsrvation deck of Yokokawa-damu
7	12R-3	12-May-12	37°35'59.0"	140° 56'21.0"	21.07	Minami-soma	Vicinity of the bsrvation deck of Yokokawa-damu
8	12K-12	5-Sep-12	37°33'23.0"	140° 52'27.7"	21.3	Minami-soma	Vicinity of the restricted area in Tetsuzan-damu
9	12N-40	22-Sep-12	37°32'27.2"	140° 51'46.4"	19.74	Namie	Roadsides (Route 49) near Haranami tunnel
10	12N-37	22-Sep-12	37°34'00.6"	140° 47'55.6"	26.01	Namie	Rear U-shaped drainageway at the milk consolidating station in Tsushima (Route 114)
11	12K-18	6-Sep-12	37°33'34.0"	140° 45'39.9"	27.96	Namie	Roadsides (Route 114) before the checkpoint from Tsushima to Namie
12	12T-20	26-Mar-12	37°38'58.7"	140° 45'20.4"	34.59	Iidate	Roadsides near Yamabe hall

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Table 2 The isotopic composition of Cs, U and Pu in Black Substances

Sample No.	Lab. code	Cs-134* (kBq/kg)	Cs-137* (kBq/kg)	U-238 (ppm)	U-236 (10 ⁻⁴ Bq/kg)	Pu-239+240 (Bq/kg)	Pu-238 (Bq/kg)
1	12N7	1408 ± 5	2337 ± 5	0.880 ± 0.004	5.21 ± 0.27	0.284 ± 0.019	0.749 ± 0.034
2	12N17	3276 ± 11	5021 ± 13	0.570 ± 0.003	1.43 ± 0.07	0.239 ± 0.026	0.552 ± 0.044
3	12N23	11410 ± 38	17670 ± 43	0.390 ± 0.009	2.40 ± 0.11	0.482 ± 0.026	1.122 ± 0.051
4	12N30	1434 ± 4	2150 ± 4	0.380 ± 0.004	1.04 ± 0.05	0.180 ± 0.012	0.411 ± 0.019
5	12R-7	2103 ± 8	2957 ± 8	0.670 ± 0.018	2.51 ± 0.16	0.273 ± 0.013	0.516 ± 0.019
6	KURO-3	4361 ± 9	5706 ± 9	1.35 ± 0.02	6.74 ± 0.57	1.14 ± 0.069	2.333 ± 0.102
7	12R-3	1809 ± 7	2492 ± 7	0.650 ± 0.013	2.83 ± 0.07	0.703 ± 0.019	1.571 ± 0.032
8	12K12	1593 ± 5	2318 ± 5	0.630 ± 0.009	2.21 ± 0.07	0.420 ± 0.015	1.001 ± 0.027
9	12N40	3282 ± 9	5053 ± 10	0.620 ± 0.004	1.56 ± 0.12	0.663 ± 0.036	1.576 ± 0.071
10	12N37	636 ± 2	947 ± 2	0.330 ± 0.001	0.45 ± 0.02	0.196 ± 0.012	0.321 ± 0.016
11	12K18	1834 ± 5	2697 ± 6	0.480 ± 0.001	0.28 ± 0.01	0.145 ± 0.005	0.252 ± 0.011
12	12T20	426 ± 3	582 ± 1	0.470 ± 0.007	0.87 ± 0.05	0.218 ± 0.021	0.508 ± 0.024

* The values are decay-corrected to the date of sampling

Error shows one sigma standard deviation from counting statistics

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Table 3 The isotopic ratios use for Cs, U and Pu in Black Substances.

Sample No.	Lab. code	U-236/U-238 (10^{-7} atom ratio)	Pu-238/Pu-239+240 (activity ratio)	Pu-239+240/Cs-137 (10^{-8} activity ratio)	U-236/Cs-137 (10^{-11} activity ratio)	U-236/Pu-239+240 (10^{-4} activity ratio)
1	12N7	2.50 ± 0.13	2.64 ± 0.19	12.1 ± 0.81	22.3 ± 1.2	18.4 ± 1.6
2	12N17	1.06 ± 0.05	2.31 ± 0.27	4.75 ± 0.52	2.85 ± 0.14	5.99 ± 0.72
3	12N23	2.60 ± 0.10	2.33 ± 0.13	2.73 ± 0.15	1.36 ± 0.06	4.98 ± 0.35
4	12N30	1.16 ± 0.05	2.28 ± 0.16	8.36 ± 0.54	4.84 ± 0.23	5.81 ± 0.46
5	12R-7	1.58 ± 0.09	1.89 ± 0.11	9.24 ± 0.45	8.49 ± 0.54	9.17 ± 0.74
6	KURO-3	2.11 ± 0.17	2.04 ± 0.15	20.0 ± 1.22	11.8 ± 1.0	5.90 ± 0.61
7	12R-3	1.84 ± 0.02	2.23 ± 0.06	28.2 ± 0.76	11.4 ± 0.3	4.03 ± 0.14
8	12K12	1.48 ± 0.04	2.38 ± 0.09	18.1 ± 0.65	9.53 ± 0.30	5.26 ± 0.25
9	12N40	1.06 ± 0.08	2.38 ± 0.11	13.1 ± 0.71	3.09 ± 0.24	2.35 ± 0.22
10	12N37	0.57 ± 0.03	1.64 ± 0.11	20.7 ± 1.28	4.75 ± 0.21	2.27 ± 0.17
11	12K18	0.25 ± 0.01	1.73 ± 0.09	5.38 ± 0.19	1.04 ± 0.04	1.96 ± 0.10
12	12T20	0.78 ± 0.05	2.33 ± 0.14	37.4 ± 3.59	15.0 ± 0.9	3.98 ± 0.45

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Table 4 The estimated core composition of Unit 1, 2 and 3 (Nishihara et al., 2013).

Unit	Pu-239+240/Cs-137	Pu-238/Cs-137	U-236/Cs-137
No.	(10^{-3} activity ratio)	(10^{-2} activity ratio)	(10^{-6} activity ratio)
1	7.86	2.29	2.87
2	7.54	1.79	2.59
3	9.96	2.29	2.71
Average	8.45	2.12	2.72

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Table 5 Summary of emission of Cs-137, Pu-239+240, U-236 and total U from FDNPP.

	Fuel composition (Bq)	Release to the atmosphere (Bq)	Release ratio (%)
Cs-137	7.0×10^{17}	1.5×10^{16}	2.10
Pu-239+240	5.9×10^{15}	2.3×10^9	3.9×10^{-5}
U-236	2.0×10^{12}	1.2×10^6	6.1×10^{-5}
Total U	6.4×10^{12}	3.9×10^6 (150 g)	

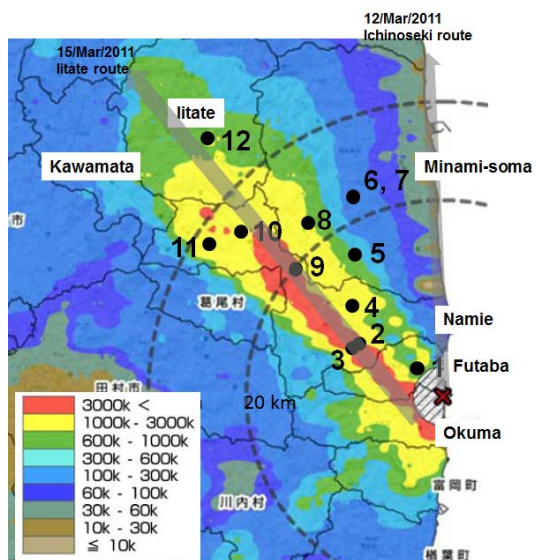
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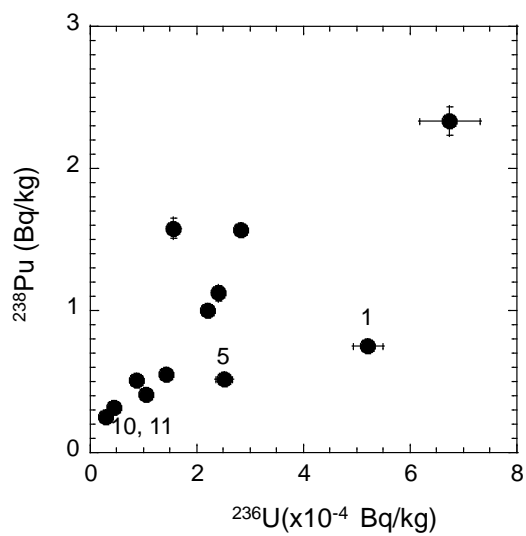
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400 Fig. 1 Sampling sites of Black Substances in Fukushima Prefecture and inventory of Cs-137
 401 (Bq/m²)⁸.

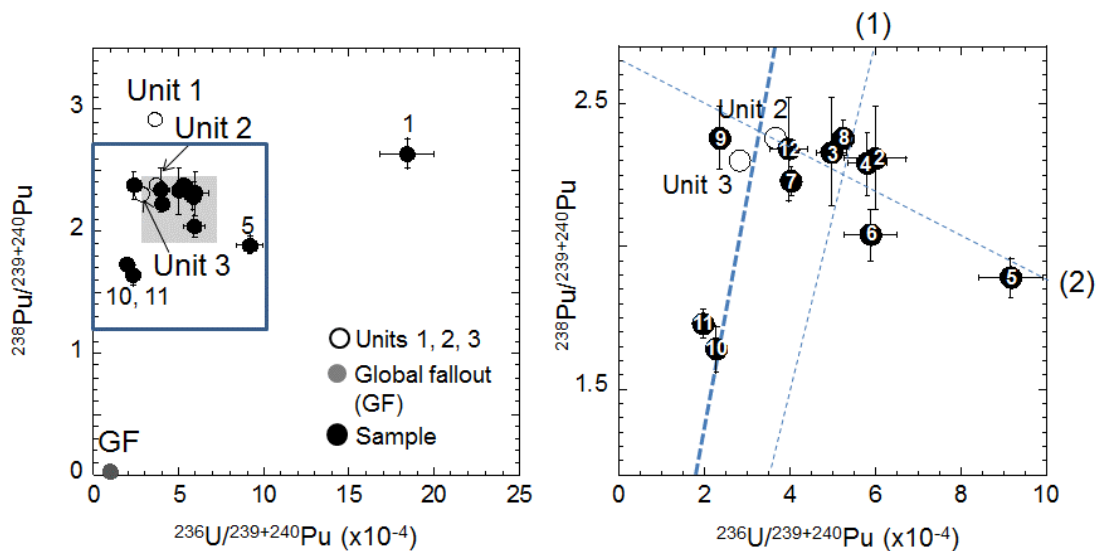
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405 Fig. 2. Comparison between ^{238}Pu (Bq/kg) and ^{236}U (10^{-4} Bq/kg) in Black Substances.

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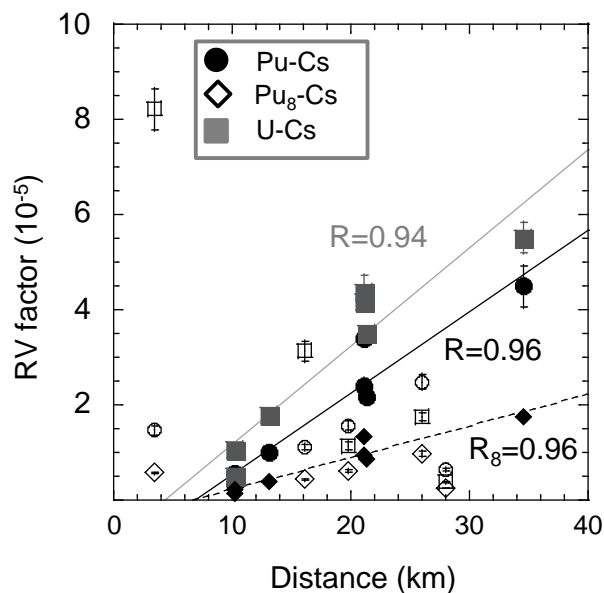


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408 Fig. 3 Relationship between $^{238}\text{Pu}/^{239+240}\text{Pu}$ and $^{236}\text{U}/^{239+240}\text{Pu}$ activity ratios obtained from Black
 409 Substances together with the isotopic composition of global fallout in Japan and the cores in
 410 each Unit (Units 1-3) of FDNPP. Grey square represent the area of one sigma (standard
 411 deviation) of $^{238}\text{Pu}/^{239+240}\text{Pu}$ and $^{236}\text{U}/^{239+240}\text{Pu}$ values which were calculated excluding sample no.
 412 1. Right graph is close-up of open square in the left graph.

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416 Fig. 4 Relationships of R_{Pu}/V , R_{Pu8}/V and R_U/V factors with distance from FDNPP to sampling
417 site (km). The solid symbols represent samples which are plotted in the grey square in Fig. 3.

418 The open symbols show other samples in Fig. 3.

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